

1 Vision of a near future: bridging the Human Health – Environment divide.
2 Toward an integrated strategy to understand mechanisms across
3 species for chemical safety assessment
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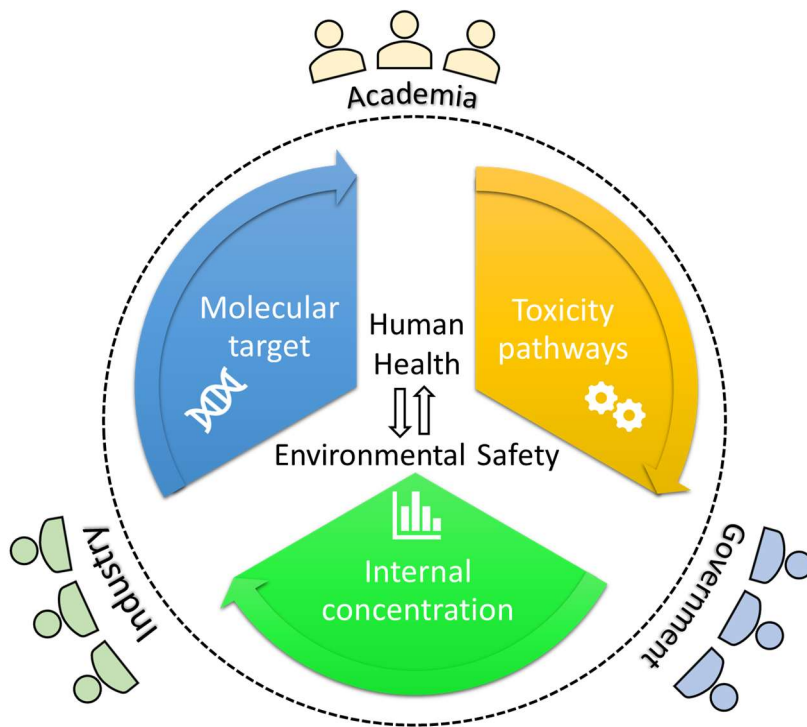
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34 Graphical Abstract



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37 1. Abstract

38 There is a growing recognition that application of mechanistic approaches to
39 understand cross-species shared molecular targets and pathway conservation in the
40 context of hazard characterization, provide significant opportunities in risk assessment
41 (RA) for both human health and environmental safety. Specifically, it has been
42 recognized that a more comprehensive and reliable understanding of similarities and
43 differences in biological pathways across a variety of species will better enable cross-
44 species extrapolation of potential adverse toxicological effects. Ultimately, this would
45 also advance the generation and use of mechanistic data for both human health and
46 environmental RA.

47 A workshop brought together representatives from industry, academia and
48 government to discuss how to improve the use of existing data, and to generate new
49 NAMs data to derive better mechanistic understanding between humans and
50 environmentally-relevant species, ultimately resulting in holistic chemical safety
51 decisions. Thanks to a thorough dialogue among all participants, key challenges,
52 current gaps and research needs were identified, and potential solutions proposed.

53 This discussion highlighted the common objective to progress toward more predictive,
54 mechanistically based, data-driven and animal-free chemical safety assessments.
55 Overall, the participants recognized that there is no single approach which would
56 provide all the answers for bridging the gap between mechanism-based human health
57 and environmental RA, but acknowledged we now have the incentive, tools and data
58 availability to address this concept, maximizing the potential for improvements in both
59 human health and environmental RA.

60 Keywords:

61 Risk Assessment; Human health; Environment; Cross-species extrapolation;
62 Mechanism of action

63

64 2. Introduction

65 It is recognised that new scientific improvements and their integration in risk
66 assessment have the potential to improve human health risk assessments by enabling
67 a mechanistic understanding of adverse effects and more accurate predictions of
68 biological responses [1]. Current regulatory-accepted approaches to assess chemical
69 safety are often based on a battery of *in vivo* methods and a limited number of
70 accepted *in silico* or *in vitro* approaches. However, performing toxicity studies for all
71 existing chemical substances using *in vivo* methods is not physically, ethically, or
72 financially possible. Chemical or biological read-across approaches are being
73 considered by industry and chemical management agencies as an alternative to
74 reduce the reliance on these highly resource-intensive *in vivo* tests. There is an urgent
75 need to improve current capabilities to perform chemical read-across and cross-
76 species extrapolation (biological read-across) through an improved mechanistic
77 understanding of the basic biology underlying toxicity and the chemistry-biology
78 interactions involved. Development of descriptors and alerts that facilitate chemical
79 grouping and a better understanding of the species hazard space (i.e. species that are
80 sensitive to certain chemical classes) would also be highly beneficial. In this respect,
81 there have been many efforts focused on the challenges involved in the development
82 of chemical read across, improving its scientific justification and supporting
83 documentation for use in both chemical hazard and RA. Chemical read-across and
84 grouping approaches have become some of the most commonly used alternative
85 approaches for data gap filling within analogue and category approaches [2]. These
86 efforts have led to a wide recognition of the scientific validity of these and its regulatory
87 acceptance and recently, ECHA has published a guidance document on how to
88 perform and document chemical read-across under REACH (Read-Across
89 Assessment Framework (RAAF) [3].

90 Over the last two decades, there has been a scientific and regulatory push towards
91 the development of novel non-animal approaches for safety assessment [4]. There is
92 a growing desire within the scientific community to achieve simpler, broader, faster
93 and importantly, more predictive risk assessment (RA). To achieve the desired
94 improvements in chemical RA, the current limitations concerning the generation,
95 integration and interpretation of newer types of data proposed for use in RA need to
96 be overcome. Recent developments in biotechnology and molecular biology have

97 given rise to New Approach Methodologies (NAMs) [5] that are greatly enhancing our
98 ability to address some of the data gaps faced in both human and environmental
99 toxicology. NAMs are a recently adopted concept to broadly refer to any non-animal
100 approach, methodology and / or technology, aimed at providing information on
101 chemical hazard and RA, including integrated approaches to testing and assessment,
102 data interpretation, and performance-based evaluation of test methods [6]. NAMs
103 open new opportunities to ensure RA is grounded in human biology rather than
104 replicating the results of a prescriptive list of animal tests. This is especially important
105 for mechanism-of-action-based RA. For instance, effect concentrations based on
106 perturbations in signalling pathways in human cells will likely be different from those
107 causing apical effects during rodent studies. The difference in species and level of
108 biological organization considered in the example suggest that results of such tests
109 cannot (and should not) be 'validated' against each other and should be compared
110 with caution [7]. In this respect, the use of molecular-based, high content data has the
111 innate potential to complement traditional human and environmental toxicology
112 approaches [8-11]. Indeed, their use could catalyse a paradigm shift to more proactive
113 pathway-based approaches, ultimately facilitating the development of *in silico-based*
114 predictive toxicology [12, 13]. Available data on endpoints supporting traditional
115 approaches to assess environmental and human safety, coupled with a growing
116 weight of *in silico* / *in vitro* biological pathways-based data raise the question: are we
117 already at a point where we can consider new types of data and incorporate them in
118 a new or augmented approach to RA?

119 For this to happen, frameworks such as the Adverse Outcome Pathway (AOP)
120 concept, which links the description of biological cascade from the insult at the
121 molecular initiating event (MIE) to the adverse outcome (i.e. AO - the apical
122 toxicological endpoint of concern), can be utilised [14-16]. In addition to the mapping
123 of data, the AOP concept also allows for qualitative evaluation of a pathway and its
124 overall reliability through a weight-of-evidence approach [17]. In some cases, for
125 example in the regulatory assessment of endocrine disruption hazard, a weight-of-
126 evidence-based approach has been advocated [18]. However, the next and ultimate
127 step required for this approach to be fully implemented in RA is the development of its
128 quantitative aspect [19, 20].

129

130 Use of cross-species extrapolation is a well-established concept for RA for
131 environmental safety (e.g. using toxicity data from a reduced number of model species
132 to represent the entire ecosystem biodiversity), but also for human health (e.g. using
133 laboratory studies from rodents to infer effects on humans). However, an improved,
134 more comprehensive and reliable extrapolation of biological pathways across species
135 would facilitate the use of already available toxicity data across human health and
136 environmental RA and allow for a more coherent and efficient characterization of
137 overall hazard [21]. Whilst the potential of molecular-based, high-content data and
138 mechanistic approaches has been recognized[5, 22], there are limited examples
139 where molecular level data have been extrapolated across species, including human,
140 to inform cross-species mechanistic understanding as part of the next-generation RA
141 of chemicals[5, 23, 24]. There is an urgent need for new approaches to classify and
142 (ideally) quantify inter-species similarities / differences based on mechanisms of
143 action. However, there are some pragmatic first steps that can be taken using
144 emerging and developing technologies (including OMICS) [25-27].

145 Motivated by these questions, a workshop was organized, entitled “Vision of a near
146 future: bridging the Human Health - environment divide. Roles of molecular and data-
147 rich approaches as part of an integrated strategy to understand mechanisms across
148 species for chemical safety assessment”, held at Colworth Science Park (Sharnbrook,
149 UK) on April 18th-19th, 2018. Representing academia, industry and government, thirty
150 experts were brought together from diverse fields, including human and environmental
151 toxicology and regulatory safety science, to foster this dialogue. The overall purpose
152 was to discuss how existing data can be better exploited and how new data can be
153 generated to improve mechanistic understanding across humans and environmentally
154 relevant species to better inform chemical safety decisions.

155 3. Workshop outline

156 Participants were selected based on their domain of expertise as well as their
157 affiliation, to ensure a broad coverage both in sense of background and areas of
158 interest as action domain. Stakeholders from universities (U. Cambridge, U. California
159 Berkeley, U. Birmingham, U. Liverpool, U. Exeter, U. Amsterdam, Brunel U.), private
160 sector i.e. industries (Unilever, Astra Zeneca), and governmental / regulatory bodies
161 (USEPA, EC-JRC and NC3Rs) were invited to discuss current problems and needs

162 concerning biological read-across and its implementation in current practices of RA.
163 New strategies and solutions were also proposed.

164 In preparation for the workshop, delegates were asked to reflect and share their
165 opinion on two key questions prior to the event:

166 (1) what are the main drivers to develop cross-species understanding of mechanisms
167 of action in the context of RA?

168 (2) what approaches / techniques do we foresee as better suited to provide scientific
169 evidence and increase confidence in cross-species extrapolation and what are the
170 main limitations?

171 All feedback received was analysed and provided the basis for a focussed discussion
172 in breakout groups. Based on their expertise and opinions shared prior to the
173 workshop, delegates were divided into three work groups, each addressing a different
174 level of biological organization: (WG1) target-level, (WG2) pathway-level and (WG3)
175 physiological-level, including exposure. Each group was asked to discuss the current
176 science and knowledge available and the scientific research needed to achieve full
177 potential as outlined in Table 1, focusing on the main challenges, benefits and hurdles.

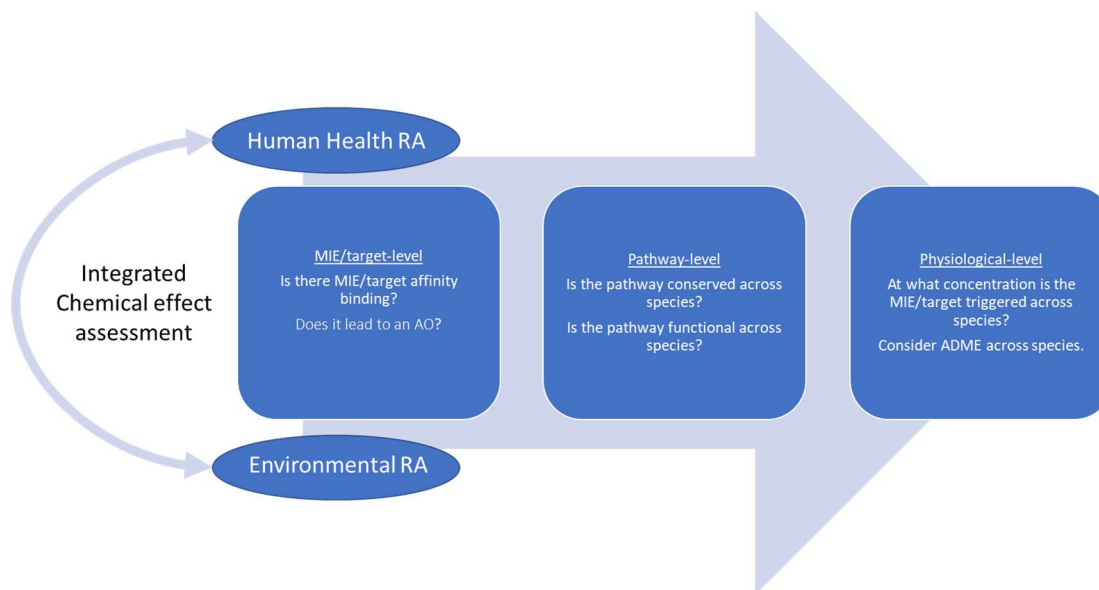
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179 **Table 1- List of the challenges considered by each of the three working groups**

Challenge 1	Improve basic knowledge of Molecular Initiating Events (MIE) across species <ul style="list-style-type: none">● Improve knowledge of target homologue/orthologue characterization through evolutionary (and functional) conservation● Develop an understanding of the chemistry of MIEs with a cross species perspective
Challenge 2	Develop basic knowledge of pathway conservation across species <ul style="list-style-type: none">● Increase scientific knowledge for pathways-based comparison to support extrapolation from a higher-tier perspective● Develop species-specific pathway-to-phenotype association analysis in a chronic exposure scenario
Challenge 3	Refine understanding of biological processes impacting internal exposure <ul style="list-style-type: none">● Understand species-specific physiological processes (ADME) to predict chemical effective doses at the target site in different species● Develop and refine PBK models for key species and link them to understand where there are common and species-specific processes

180

181 This report describes the proceedings of the workshop and presents the highlights of
182 the discussion. All opinions were treated equally and were consensually accepted by
183 all participants.



184
185 *Figure 1-Holistic schematic of how to perform chemical risk assessment using mechanistic knowledge improving the cross-*
186 *talk between human health and environmental safety.*

187 4. Breakout Discussion group summaries

188 4.1 Work group 1: “The Challenge: Improve our knowledge of MIE across species”

189 Key shortcomings regarding the use of MIEs (the initial chemical–biological interaction
190 that starts the AOP [28]) to identify and understand common pathway signal
191 transduction for cross-species extrapolation were considered by WG1. When
192 considering a MIE for RA across human health and the environment, it is important
193 first to recognize that there are different aims and diverging protection goals. In fact,
194 for humans the protection goal is optimally set at the individual level, aiming at
195 protecting each individual against harm; for the environment, this is more often
196 established at the population or ecosystem level [29]. Similarly, there may be varying
197 layers of complexity to consider, for example general narcosis vs. specific Mode of
198 Action (MoA), or a MIE with multiple interactions (e.g. skin sensitizers) vs. a MIE that
199 leads to one specific adverse event (e.g. estrogenic receptor agonist). Challenges also
200 differ substantially depending on the goal for extrapolation between datasets / species,
201 i.e. to assess for a similar MoA / AOP (or toxicity pathway) across species or to
202 extrapolate effect levels across species. If the aim is to assess for similar MoA / AOP
203 across species (but not effect levels), current state of the art ortholog predictions (e.g.

204 OrthoDB [30] or EggNOG [31]) can provide a good starting point, provided the
205 mechanism of toxicity is specific and the MIE limited to one (few) specific protein
206 targets. However, there are many uncertainties associated with ortholog predictions.
207 For example, uncertainty increases with evolutionary distance between species as
208 well as for some types of protein families such as CYP450 or G-proteins [32]. In the
209 pharmaceutical field this has been addressed by using a majority vote across three
210 prediction platforms in a web-based application that looks for protein target
211 conservation between human and a range of sequenced phyla (www.ecodrug.org)
212 [33]. Furthermore to this, another tool facilitates summary, comparison and access to
213 various sources of ortholog predictions and provides a comparison of 17 different tools
214 and algorithms to increase the confidence in the orthologue prediction
215 (<http://www.flyrnai.org/diopt>) [34]

216 To understand how a MIE could be used to inform RA and enable cross-species
217 extrapolation, the extent of functional conservation of downstream effects across
218 species also needs to be resolved. This could be achieved by deploying new functional
219 *in vitro* assays, although this is expected to be time and resource intensive. However,
220 some compounds will interact with multiple targets, and may lead to different
221 downstream events. This highlights the importance of understanding the response of
222 biological systems from a network perspective [20, 35]. Moreover, since chemical-
223 target(s) interaction networks are often driven by internal exposure dynamics, it is also
224 essential to enhance the understanding of adsorption, distribution, metabolism and
225 excretion (ADME) processes, especially in lower species, where current knowledge is
226 limited. This will allow enhanced consideration of chemical-target interaction networks
227 that may occur following diverse exposure scenarios, thus simplifying and boosting
228 the cross-species extrapolation process.

229 Furthermore, it was agreed that the pathway leading to the AO itself needs to be fully
230 understood to prioritize testing needs. Understanding how gene / proteins relate to
231 downstream functions through evolutionary relationships between protein families and
232 super-families may also be informative and more meaningful than a one-to-one
233 comparison. Therefore, it is important to understand the available data including
234 substrate specificity and related potencies to discern how the level of gene / protein
235 similarity influences the target affinity and the impact on potency.

236 While doing so, it is important to keep in mind the whole decision-making process, to
237 better and more efficiently define what is the minimum but necessary information
238 required to enable decision-making for RA purposes and thus reduce overall
239 uncertainty. Pertinent to this, it will also be paramount to consider the different needs
240 of different stakeholders (e.g. regulators vs. industry).

241 The need for a deeper understanding of the difference between receptor-mediated
242 and more general stress responses was also discussed. Potential solutions included
243 the idea of developing directed functional bioassays, as well as building a library of
244 target-knockout systems encompassing several species. To ensure meaningful
245 results and application, any of these approaches would need to make use of a broad
246 selection of chemicals, representing a variety of chemical classes and MoAs / AOPs,
247 as well as to cover different suitable exposure durations and time points (including life
248 stages), ensuring coverage of potential sources of variability. Ultimately, these
249 approaches would generate a repository which could then be interrogated for hazard
250 characterization every time a new substance comes in for hazard evaluation. There
251 are already ongoing efforts pointing to this same direction, including the Library of
252 Integrated Network-Based Cellular Signatures (LINCS) Program
253 (<http://www.lincsproject.org/>) [36], providing a first attempt to create a network-based
254 library of biological signatures by cataloguing changes in gene expression and other
255 cellular processes occurring when cells are exposed to a variety of perturbing agents.
256 While this represents a powerful source of information, it is currently limited to human
257 and more of these kinds of approaches are needed to support the evidence across
258 species. Nevertheless, it is recognised that chemical exposure levels in the
259 environment are often very low at a cellular level and producing assays with
260 environmentally relevant cellular exposure becomes difficult (reiterating the need for
261 cell level exposure considerations). Another discussed alternative, and potentially
262 more efficient, way to test this concept would be to start using available data,
263 comparing current existing human toxicity signatures (for instance, from the LINCS
264 database, among others) to available historical toxicity records in the ecotoxicology
265 literature. The proposed database would be used as a surrogate to define the
266 biological target space and could be interrogated to identify potential consensus
267 hazard signatures, based on effect conservation. However, this approach would also
268 pose several other practical questions: what data types would that database include,

269 such as life-history, transcriptomics, metabolomics, etc.? On which species and
270 chemicals? How would it be prioritized? All of these questions highlighted the recurrent
271 need to increase the ecological realism by considering a larger number of species and
272 thus, related delivered ecosystem functions. This also implies that thorough
273 predictions of pathway-based signatures are urgently needed to better estimate risk,
274 especially when trying to define the most relevant / appropriate species under each
275 scenario. Increasing the two-way data flow (human health to ecotoxicology and vice-
276 versa) would undoubtedly improve the understanding in this field.

277 The final note from the Work group 1 discussion was the recognition that
278 environmental RA information is not currently being fully exploited within the human
279 health arena (and vice-versa). In fact, there is still a great potential for developing
280 additional biological read-across and extrapolation processes from human health to
281 environmental safety science (and vice-versa), but their different needs and priorities
282 need to be acknowledged. In this sense, generation of new data may not be a priority
283 need, but rather the development of new / improved data mining tools to interrogate
284 the wealth of data that is already available.

285 4.2 Work group 2: “The Challenge: Develop our basic knowledge of pathway 286 conservation cross-species”

287 This group addressed how to tackle cross-species extrapolation at a pathway-level
288 and discussed several key issues that need to be resolved to increase confidence
289 before application. It was identified that gene function is the crucial aspect in this
290 respect and the concept of “functional orthology” [37] was considered a beneficial
291 approach to predict the conservation of the (adverse) outcome across species. To
292 address and expand this concept, investigations on different levels are needed.
293 (Re)defining and cataloguing the orthologs by function would help sorting and
294 functionally annotating them into the relevant pathways. There is high probability that
295 a number of genes and gene subfamilies are divergent across species, and
296 additionally, multi-purpose enzymes found in lower species may replace their role and
297 thus belong to multiple pathways. A first attempt in this same direction is provided by
298 the new available software Gene2Function (<http://www.gene2function.org/>) [38]
299 whose primary goal is to facilitate the development of new hypotheses regarding the
300 function of a given gene based on what is known about the function of orthologs of
301 that gene in other species.

302 Going beyond a better functional annotation, before stepping-up to a purely pathway-
303 level analysis, the need for a more human and environmental toxicology-relevant gene
304 annotation was also acknowledged. In fact, essential genes / gene families involved
305 in human health (inferred by the many medical / pharmacological studies available)
306 may not always be the same as those that are of ecotoxicological concern. Therefore,
307 as a possible solution, it was suggested to map the human genome against existing
308 ecotoxicology literature in a newly designed, fit-for-purpose database, thus re-
309 annotating genes based on ecotoxicological needs.

310 The participants acknowledged attempts to define all known AOPs in human and
311 environmental toxicology, however the data are currently far from complete and little
312 is known about cross-species evaluation. The need to define a priority list of the most
313 relevant pathways was discussed and agreed it would provide a good starting point
314 for deeper exploration. One proposed hypothesis was that the “key” pathways that are
315 essential for life are likely to be the more evolutionarily conserved. These could include
316 pathways such as oxidative stress, Nrf2, the p53 DNA damage response, the unfolded
317 protein response (UPR) and mitochondrial injury, among others [39]. Exploration here
318 should be focused both on improving understanding on both an evolutionary scale and
319 on an experimental level. For example, it was suggested the creation of a priority
320 pathways screening panel across relevant species, including new *in vitro* assays for
321 toxicity and stress responses coupled with Physiologically based Kinetic (PBK)
322 models. However, improved insights on the level of pathway conservation is required
323 to be able to interrogate their (potential) de-regulation to the initiation of apical effects
324 (or the lack thereof). This would also serve to improve the functional annotation of the
325 pathways themselves, as mentioned above, thus developing a new “apical functional
326 ontology”. Though, even if some of the pathways are conserved between species, the
327 apical endpoint might not be present, could be organ-specific or could manifest itself
328 in a different (not directly identifiable) way. AOP-Wiki (<https://aopwiki.org>), the central
329 repository for all AOPs developed so far, represents a good source of information to
330 identify the known links between MIEs, and the cascade of key events (KE) leading to
331 the apical endpoints / AO [15]. A good example of the former is AOP 150 “Aryl
332 hydrocarbon receptor activation leading to early life stage mortality, via reduced
333 VEGF” where the developing embryos of birds and fishes are most sensitive to the
334 stressors activating this AOP, ultimately leading to embryo death and population

335 trajectory decline; mammals appear to be less sensitive, leaning towards cardiotoxicity
336 that persists into adulthood, and increasing susceptibility to heart disease rather than
337 embryo-lethality. It was also discussed that sub-pathway modules / key events might
338 be more conserved, thus easier to track, and might give more information between
339 different species than investigating the whole pathway. While it is appreciated that we
340 cannot expect to unravel all AOs for all toxicity pathways of concern in all species,
341 moving toward these kinds of approaches would help to increase confidence in toxicity
342 predictions. Also, it would be a significant advancement to know when a pathway is
343 conserved and disrupted across which species, thus defining with more confidence
344 the space for environmental risk.

345 In this respect, it was acknowledged that the aim of research is always to advance
346 science to serve society with the maximum knowledge possible. As scientists, it is
347 appreciated that curiosity drives the understanding of every mechanism and
348 interaction between a chemical and an organism. Regulatory pressures require data
349 underpinning human health or ecological assessment to be consistent and robust, with
350 the goal of ensuring safety to humans and the environment. This raises the challenge:
351 when do we have enough data for RA? An admittedly complex question to which there
352 is no easy answer. The ultimate goal is to achieve enough confidence to enable
353 decision-making without endless laboratory testing and years of research. Starting
354 from the point that it is not realistic to cover all aspects for assessment contexts, it was
355 suggested that it may be easier to know when the data is not enough. It is a matter of
356 increasing confidence and reducing uncertainty: for instance, one could hypothesize
357 that enough data might mean having several pathways annotated to allow satisfactory
358 toxicity predictions, although we don't know all of them. Thus, to answer the question
359 on how much investment is acceptable to reduce the uncertainty of risk, we first need
360 to think of how current uncertainties are preventing decisions to be made and how
361 much we are willing to invest (in time, money and effort) to improve this certainty. The
362 current revolution in digital technology and machine learning approaches may well
363 help to address both the question of how much data is required, and what information
364 is retrievable from existing data [40] .

365 From an environmental RA standpoint, it will never be possible to consider all
366 environmental species and all possible exposure scenarios (time, doses, frequency
367 and ecological circumstances). As such, there is a need to derive sufficient evidence

368 allowing to build models and provide enough scientific basis to support reasonable
369 predictions based on relatively small datasets. There is not one unique environmental
370 relevant species that is better, more representative or more appropriate than others.
371 A concept currently being explored by evolutionary biologists considers several
372 species (5-8 species) covering the phylogenetic tree in its main branches. This could
373 be considered the minimum number of species needed to reflect the main
374 distinctiveness of evolution. In any case, it continues to be very challenging to include
375 species-specific physiology into the equation as well as the ecological traits that are
376 unique to each (sub-)species. Similarly, it is important to understand how to consider
377 and account for genes that may exert different roles simultaneously, or different
378 functions throughout the lifetime of the organism; how to deal with epigenomics and
379 the knowledge that genomes are adaptive to environmental conditions and / or
380 external stimuli; how to overcome the potential problem that genes may behave
381 differently when tested in controlled lab conditions as compared to their native state.
382 All of these remain open questions: although much work has been done to try to
383 answer these questions, comprehensively addressing these and other concerns on a
384 case-by-case basis are still far from application and beyond the current requirements
385 for RA.

386 4.3 Work group 3: “The Challenge: Refine our understanding of biological processes 387 impacting internal exposure”

388 Consideration of the main biological processes impacting internal exposure, and
389 particularly, how species-specific ADME processes influence chemical concentrations
390 at target sites, is critical to the application of mechanistically-based species
391 extrapolation. The concept of “exposure” cannot usefully be discussed in isolation, but
392 rather as an integrated part of the RA question. Without inclusion of exposure, any
393 discussions on chemical-target interactions and species extrapolation of the
394 responses remain theoretical, limited purely to the identification of similar hazards. In
395 order to translate an identified hazard into a risk, considering and understanding
396 exposure is essential. Being RA is driven by exposure, it is not relevant that the
397 molecular target triggered by a given compound is conserved across species, if the
398 exposure level is below the activation threshold of the MIE. Thus, it becomes essential
399 to understand the RA question, and link to the specific exposure scenario, and the
400 required level of confidence needed to make an early decision on risk. At sufficiently

401 high chemical exposure doses, organisms and cells often exhibit acute effects related
402 to general membrane perturbation, e.g. narcosis, whereas at continuous but lower
403 doses, different pathways may trigger measurable effects at various thresholds. In
404 addition, complexity is also added by differences in sensitivity between different cell
405 types as well as different rates of metabolism across tissues and organisms. Indeed,
406 when extrapolating from *in vitro* to *in vivo*, or from species to other species, differences
407 in both biokinetic and biodynamic properties are of central importance and neither can
408 answer the question of risk independently.

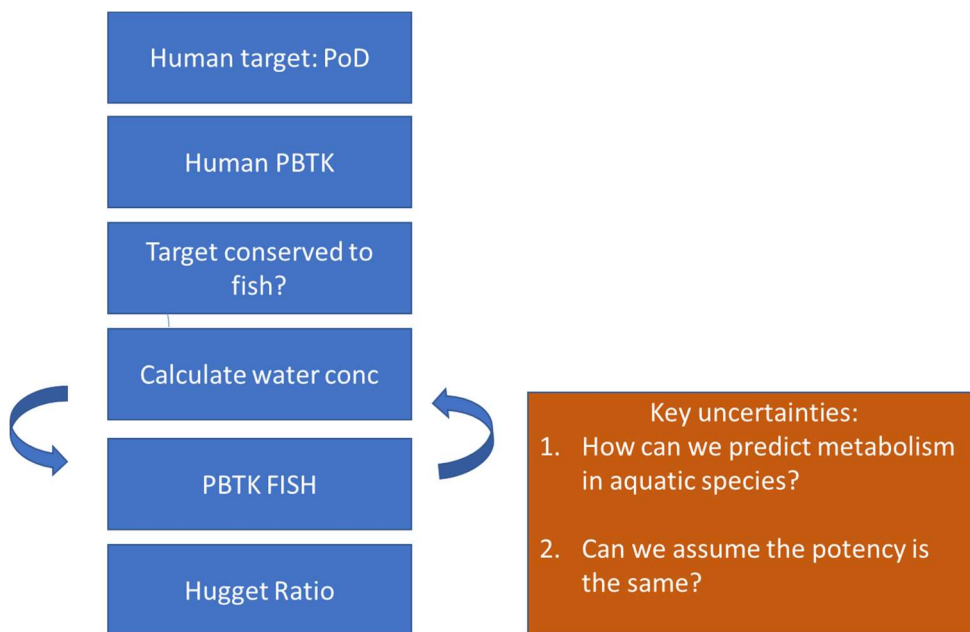
409 Lessons can be learned from human health where for decades animal effect data have
410 been utilised to extrapolate to potential human effects. More recently, research has
411 been focused on extrapolating from *in vitro* to *in vivo* effects to eliminate and overcome
412 the need for animal testing. The magnitude of the challenge of applying the same
413 strategies to environmental RA is apparent: rather than dealing with one very well
414 characterised organism, thousands of diverse, highly variable and poorly
415 characterised organisms need to be considered. Given this complexity, the workgroup
416 focussed discussion mostly around fish, where extrapolation approaches from existing
417 data is key, given the desire to eliminate animal (i.e. vertebrate) testing.

418 In order to meet these needs, the MERLIN-Expo software (<https://merlin-expo.eu/>)
419 was developed, which contains a library of models for exposure assessment coupling
420 environmental multimedia and pharmacokinetic models, and aims to link
421 environmental fate of chemicals and internal concentrations in humans, thus
422 integrating environmental exposure assessment and human exposure assessment.
423 Although it represents a very significant step in this space, it is centred over human
424 health RA and does not cover the heterogeneity found in the environment in terms of
425 species and ecosystems, that still need to be addressed further for its implementation
426 in ERA.

427 Across the pharmaceutical industry as well as pesticides and biocides, chemicals of
428 interest are designed for high levels of specificity and potency, and effective
429 absorption. This combination of chemical attributes can often lead to measurable
430 effects at realistic exposure scenarios for aquatic species despite being designed for
431 low bioaccumulation potential [41]. As such, there has been a pressing need for a
432 common strategy for environmental RA for these industries. In contrast, ingredients

433 used in Home and Personal Care (HPC) products are designed to be of low bioactivity
434 as possible. As such, toxicological concern associated with these types of chemicals
435 is reduced, though the volumes used are greater compared to pharmaceuticals, for
436 example.

437 It is therefore understandable that the greatest examples forward in terms of MIE
438 identification and species extrapolation come from the pharmaceutical sector. One of
439 the most prominent strategies for environmental RA coming from this sector is the
440 Hugget approach (Fig 1) [42-45]. This method presents an approach for biological
441 read-across from human therapeutic doses to environmental species (e.g., fish). It
442 rather simplistically compares internal concentrations in fish and human based on
443 toxico-kinetic modelling and environmental fate calculations.



444

445 Figure 2 Hugget approach scheme. Acronyms PoD: Point of Departure, PBTK:
446 Physiologically-based ToxicoKinetics

447 However, the application of this approach in other sectors is more challenging
448 because, for instance, HPC ingredients rarely have a full package of ADME data
449 associated with them, predominantly because they were not designed for biological
450 interaction in the same way that pharmaceuticals are. This could suggest that for many
451 chemicals the Hugget approach may be excessive due to its extensive data
452 requirements. However, the exploitation of existing data from other compounds or from
453 other species may obviate the need for extensive testing if the uncertainties
454 surrounding these data can be addressed. These uncertainties result from 1) whether

455 we can assume metabolic machinery is significantly similar across species to predict
456 metabolism in relevant species and 2) whether the potency of target effects is the
457 same in different species.

458 Given the relative abundance of data generated regarding human metabolism, a
459 consideration to be made regards how acceptable it is to read across from data
460 generated to satisfy human safety needs (using *in vitro* assays) to fish and other
461 species. This raises a number of research questions: a) Can we cover all the biological
462 space for fish using existing human cell lines? b) Can we use existing cell line data
463 sources (e.g. American Tissue Cell Collection or Cellosaurs) to define suitable cell
464 lines to cover that biological space? c) have we performed extensive comparisons
465 between these human-fish cell lines and if required can we establish new lines to fill
466 the gaps? d) Can we use existing untargeted chemistry data e.g. a metabolomics
467 study, to see and / or model metabolism patterns for selected chemical classes across
468 human and model species? Can we take broader approaches for understanding
469 metabolism in environmental species? For example, can we classify clearance in fish
470 as high-medium-low and can we establish uptake rates for chemical classes? In this
471 regard, the workgroup also discussed the possibility to perform computational
472 simulations to address the uncertainty across a range of species and use randomness
473 to generate draws from a probability distribution. Main strength of this approach is the
474 possibility of considering any potential outcome of a process and thus, assessing the
475 whole impact of risk and allowing for better decision making under uncertainty and
476 deficiency of data.

477 When looking into the application of these approaches for RA, it was acknowledged
478 that additional careful consideration is needed in designing a new exposure-driven
479 environmental RA strategy. The conventional Predicted Environmental Concentration
480 / Predicted No Effect Concentration (PEC / PNEC) strategy for assessing
481 environmental risks is often criticised for its lack of environmental realism and its
482 conservative nature. The strategies adopted by the pharmaceutical sector require
483 extensive data generation that may not always be feasible for other products or
484 ingredients. However, a strategy based on maximizing read-across data from other
485 species principally focussing on understanding and characterising metabolism may
486 negate the need for large-scale metabolism data generation. Additionally, this should
487 enable to maximize the value of existing PBK approaches to establish internal

488 exposure concentrations as part of a broad modelling approach for environmental RA.
489 Gaps in knowledge still exist in order to determine which approach is most appropriate
490 and more species-specific PBK models and metabolism data is required to be
491 generated to support. Overall, the group concluded that there is a need to develop
492 further approaches to capitalize on the advances made in molecular target discovery
493 and being able to determine internal exposure to a sufficient level of accuracy, whilst
494 still maintaining a pragmatic approach.

495 5. Discussion and future outcomes

496 Over the last decade, there has been advancement in the way that chemical RA is
497 performed and there has been an accelerating global shift toward animal-free
498 methods. The International Cooperation on Cosmetics Regulation (ICCR) recently
499 defined Application of 'Next Generation RA', as an exposure-led, hypothesis-driven
500 RA approach that integrates *in silico*, *in chemico* and *in vitro* approaches, and provides
501 an example of how this framework is becoming more embedded [7]. Nevertheless,
502 current environmental RA standard regulation guidelines still rely on extrapolating
503 largely *in vivo* data from a limited number of model species to a multitude of species
504 of environmental concern using safety factors to account for uncertainties. This
505 protective rather than predictive (or realistic) approach to biological read-across
506 presents significant barriers to the broader use of the increasing wealth of NAMs-
507 derived data for inferring impacts across organisms within a Next Generation RA
508 framework. At present, biological read-across tends to be constrained by its large
509 degree of uncertainty due to inherent physiological diversity (obvious at the organ to
510 species-level, but less obvious at the sub-cellular MIE-level), the wide range of
511 sensitivities to chemicals and the limited mechanistic understanding of toxicity in non-
512 target species. In fact, the general lack of comparative cross-species sensitivity data
513 (including human) limits the ability to make robust taxonomic extrapolations in support
514 of RA. Overall, it is essential to increase trust in these methods, by building confidence
515 among regulators and the broader scientific community that the necessary biology is
516 comprehensively and adequately incorporated into the proposed animal-free
517 strategies, so that they can be applied and used for both human health and
518 environmental RA. To encourage this process and to reduce the associated
519 uncertainty, it is imperative to identify and explain the relevant inter-species similarities

520 / differences, to allow more evidence-based extrapolations and an improved
521 assessment of uncertainty.

522 There are several ways to achieve this. A common concept is based on pure orthology,
523 which relies heavily on sequence similarity and phylogenetic events, and is illustrated
524 by SeqAPASS (Sequence Alignment to Predict Across Species Susceptibility,
525 <https://seqapass.epa.gov/>) [46]. This tool attempts to answer the question whether or
526 not a known protein target is present in another species for a chemical to act upon.
527 Information from SeqAPASS, in concert with AOP descriptions can begin to inform the
528 potential for cross-species effects propagating from a MIE to an AO. Another approach
529 "Interspecies Correlation Estimates" (<https://www3.epa.gov/webice/>) [47], was
530 developed by the USEPA with two aims: the estimation of acute toxicity from a
531 surrogate species, and a species sensitivity distribution model which generates a
532 prescribed hazard level. The main disadvantage of this approach is that it is based on
533 statistical inference and modelling and is not inclusive of any biological or mechanistic
534 information. While acknowledging the added value of these approaches in providing
535 additional lines of evidence applicable to a WOE evaluation in a decision-making
536 context, they presently still lack the underlying mechanistic understanding needed to
537 improve safety decisions. Yet, the tools are not static and continue to evolve as the
538 science advances in this area of bioinformatics.

539 The workshop discussed the opportunity for application and the limitations of the
540 current approaches, along with proposing NAMs that could improve biological read-
541 across while reducing the inherent uncertainty embedded in the process. At the same
542 time, it would allow greater re-use of existing data sources. In this regard the main
543 outcomes of the workshop can be distilled into an augmented concept of "functional
544 orthology", in which the common orthology concept should be merged with functional
545 and mechanistic information, namely the information being generated by ToxCast,
546 SEURAT-1, EU-ToxRisk projects among others [48-51], thus expanding the
547 understanding of the underlying processes leading to toxicity. Although this is not a
548 completely new concept, it is only now that the latest advances in technology and
549 knowledge will allow "functional ontology" to be fully achieved. In this perspective, the
550 most effective and unbiased way to determine gene function is through functional
551 genomic studies, which usually involves (systematic) knockout of genes, followed by
552 assessment of phenotypes such as lethality / viability, growth, development, etc.

553 These “genotype-to-phenotype” approaches can also provide insight into chemical
554 mechanisms of action, helping to define more specific toxicological endpoints and
555 informing the development of novel mechanistic-based toxicity bioassays. New
556 powerful molecular techniques are now available to obtain targeted knockouts, such
557 as homologous recombination, RNA interference (including siRNA and shRNA),
558 engineered site-specific nucleases (i.e. zinc-fingers, TALEN, CRISPR). However,
559 using only sequence homology (and / or orthology) as a basis for extrapolation may
560 be somewhat limited (as discussed previously) and not a guarantee of the
561 conservation of function. Thus, managing to overcome this assumption by considering
562 the functional level would lower the ambiguity and reduce overall uncertainty. The
563 translation of these new understanding into novel pathway maps could be used to
564 better define the species-impact space for well-defined toxicity pathways. This is not
565 a trivial task; though, if collective and coordinated efforts are brought forward, ensuring
566 gaps are addressed a by all the relevant stakeholders and new knowledge is then
567 translated into regulatory changes, the benefits for both human health and
568 environmental RA are expected to be highly significant.

569 Since most of the currently generated NAMs data are designed to inform human health
570 assessment, ecotoxicology has much to gain from an increased knowledge of pathway
571 homology. There is high potential to benefit from a comprehensive and well
572 understood mechanistic-based predictive science, addressing the long-standing issue
573 of chronic (i.e. long term) sub-lethal exposure. Nevertheless, there are also potential
574 advantages for human health RA, such as the concept of new PBK and dynamic
575 models developed for chosen invertebrate species that may be the missing piece of
576 the puzzle and provide the evidence of whole-organism function. This will offer human
577 health researchers a new multi-dimension, fully functional biological model, helping
578 better inform the processes involved in toxicity and / or disease. Human health
579 researchers have recently developed several approaches to overcome the
580 shortcomings of single cell line testing (e.g. lack of biological relevance, impaired
581 metabolism) by using multicell plates, organotypic 3D models, among others. Although
582 a big step forward and more relevant *in vivo* conditions, these approaches may not yet
583 be sufficiently representative of the dynamics of a whole organism. In this sense, the
584 development of a deeper biological / physiological knowledge of invertebrate species
585 and how they deal with stress, can have a large impact. Existing vertebrate-based

586 PBK models can already provide valuable information on internal concentrations
587 (particularly in fish) to support RA decisions, although broader applicability of such
588 models may be obtained through closer investigation of data read-across potential for
589 some input values and the associated uncertainties. For instance, until recently, when
590 considering the traditional protection goal of environmental RA as the population level,
591 decision-making, , has largely been based on the test chemical concentration able to
592 disrupt apical processes, (i.e. development, growth, reproduction, and survival) in
593 environmental species, which may ultimately alter population dynamics. In this regard,
594 the issue of estrogenic chemicals in the aquatic environment provides an interesting
595 example of the practical significance of this aspect. The intersex condition (i.e.
596 presence of eggs in the testis) observed in male fish in the rivers of many western
597 countries is probably one of the most dramatic phenotypic effects observed in
598 freshwater wildlife associated with chemical contamination [52, 53]. The discovery of
599 intersex in wild fish, in the 1990s, triggered an entirely new stream of research,
600 endocrine disruption in the aquatic environment. The evidence produced by this large
601 volume of research led, in 2015, to the inclusion of 17-Alpha-ethinylestradiol (EE2),
602 17-Beta-estradiol (E2), and estrone (E1) in the surface water Watch List under the
603 Water Framework Directive. However, concurrently, a large study carried out in the
604 UK demonstrated that populations of cyprinid fish are self-sustaining despite
605 widespread feminization of males [54], raising a new challenge for the interpretation
606 of the relevance of testicular intersex for decision-making in environmental RA. This
607 suggests that more efforts should also be dedicated to the evaluation of whether
608 implementation of the latest mechanism-based predictive toxicology approaches
609 would bring significant benefits compared to the traditional approach. Regardless the
610 aim of the RA, predicting adverse phenotypes triggered by exposure to chemicals
611 remains an essential aim of modern toxicology. This challenge is particularly pressing
612 if we consider that millions of animals would be required to test the potential toxicity of
613 the thousands of chemicals currently in commerce. The application of the AOP
614 concept to frame existing toxicological knowledge from a mechanistic perspective has
615 been proven to provide a good platform to bridge the gap between human and
616 environmental RA [20], and to support the regulatory acceptance of mechanistic
617 considerations in an environmental RA context.

618 6. Conclusion and recommendations:

619 Overall, the workshop highlighted a clear and common motivation to progress towards
620 the application of mechanistic-based animal-free chemical safety assessment
621 methods. However, there is no single fit-for-all approach and it is clear that seldom will
622 be possible to directly replace animal testing with NAMs, but rather coordinated efforts
623 aimed at an integrative implementation of new approaches are required. As a first step,
624 it was urged further exploitation and integration of the wealth of already available
625 information. Better employment of existing data and tools may drive toward an
626 improved cross-species extrapolation and lead to reduced reliance on animal data.
627 This is particularly true for environmental toxicology where similar tests are traditionally
628 required in multiple species to meet global regulatory requirements, but also useful to
629 bridge the existing knowledge gap between human and environmental toxicology.
630 Moreover, further targeted development of NAMs and generation of ad-hoc data would
631 greatly increase confidence and scientific evidence for extrapolating MIEs, KEs, or
632 entire pathways (e.g. MoAs / AOPs) between human and environmental relevant
633 species, thus consolidating the shift to a more mechanistic-based predictive RA and
634 support the use of a broader landscape of data across both human health and
635 environmental RA fields.

636 The identified and prioritized research needs and key recommendations from the
637 workshop cover both current technical challenges (i.e. required research & capability-
638 build) and decision-making challenges (i.e. development & evaluation for RA), as
639 follows:

- 640 ● Research needs:
 - 641 ○ To setup collaborations between relevant stakeholders (academia, industry,
642 regulators, NGOs) to define endpoints or AOs of concern across human models
643 and environmental species, leading to prioritized testing needs.
 - 644 ○ Define a priority list of pathways of environmental toxicological concern, which
645 are key to organism survival, growth and / or reproduction and describe the
646 extent of their conservation across species.

- 647 ○ Build a database designed around functional gene annotation to enable full
648 exploitation of data from all relevant species coupled with improved mining tools
649 to adequately interrogate this data.
- 650 ○ Address uncertainties in orthology assignment by re-designing and deploying
651 functional assays to report downstream / upstream pathway-based effects.
- 652 ○ To develop cross-species relevant screening panels for the identification of
653 priority pathways able to predict AOs of concern.
- 654 ○ To develop approaches linking the environmental / external concentration to the
655 cellular / internal concentration at the target tissue and the AO in order to
656 understand the minimum level of perturbation necessary to trigger toxicity.
- 657 ○ Improve the understanding of ADME processes across key relevant biological
658 classifications (e.g. family or species).
- 659 ○ Develop computational tools (i.e. models) to enable prediction or classification
660 of chemical clearance rates for species relevant to RA, namely simplified PBK
661 models.
- 662 ● Decision-making challenges:
 - 663 ○ Identify current technology and application gaps which need to be addressed
664 for the successful implementation of NAMs in RA.
 - 665 ○ Establish and promote confidence in extrapolating effects across species using
666 mechanistic data, through the development of case-studies
 - 667 ○ Develop a new globally harmonized RA framework incorporating NAMs and
668 making use of all the data available (using well established chemical read-
669 across as well as cross species extrapolation where possible)
 - 670 ○ Improve the understanding of exposure scenarios, namely reducing its
671 granularity, so that those considerations can be better incorporated into RA.
 - 672 ○ Ensure early engagement and maximize communication of the private sector
673 (i.e. industry) and regulatory agencies to drive and ensure future fit-for-purpose
674 in the scientific / technical development of NAMs.

675 As a last note from the workshop, it was remarked that it is imperative to continue the
676 flow of these discussions to include the wider scientific community, regulators, and

677 industry, while continuing to progress the development of novel scientific approaches
678 to fully explore the potential of NAMs. All relevant stakeholders should be involved in
679 this discussion to ensure its proper development. Only by having developers and end-
680 users discussing the advancement of new approaches together we can ensure that
681 they are fit-for-purpose and meet the innovation and decision-makers' needs.

682 Disclaimer:

683 All authors are employees of their respective organisations. Their views expressed in
684 this manuscript are their own and do not necessarily represent those of their
685 institutions or companies.

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